

Red River Floodway Expansion Project 2014 Post-Construction Monitoring Program Activity Report Memo Reference: .9999-93.07 HM100

DRAFT - Rev A

KGS Group 05-1100-01.19.12.06 October 2014

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October 10, 2014

File No: 05-1100-01.19.12.06

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ATTENTION: Ms. Leanne Shewchuk

Manager Special Projects and Environmental Services

RE: Red River Floodway Expansion Project

2014 Post Construction Monitoring Program Activity Report

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Draft - Rev A, October 2014

Dear Ms. Shewchuk:

Please find enclosed one (1) copy of the draft 2014 Post-Construction Monitoring Program Activity Report. The report satisfies requirements for annual monitoring in Environmental Licence 2691 for the 2014 period, and was conducted according to the Post Construction and Long Term Monitoring Plan HM72 Rev 1 issued April 2013.

We appreciate the opportunity to provide ongoing services to the Manitoba Floodway Authority.

Sincerely,

J. Bert Smith, P.Eng. Channel Design Manager

JBS/MFH/mlb Enclosure

Cc: Dave MacMillan - KGS Group

Marci Friedman Hamm - KGS Group

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HM100-1 2014 Monitoring Locations

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1.0 INTRODUCTION AND AQUIFER CHARACTERIZATION

This 2014 Groundwater Monitoring Activity report is submitted in response to the requirements for annual monitoring in accordance with Clause 27 and Clause 30 of Environmental Licence No. 2691 dated July 8, 2005 and described in memorandum HM72 Rev 1 Post-Construction and Long-Term Monitoring Program issued April 2013. Groundwater activities for 2005 and 2006 were summarized in the 2006 Groundwater Monitoring Activity Report issued March 2007, which should be used as a reference to this report. Activities for 2007 to 2011 are summarized in the annual 2007 to 2013 Groundwater Monitoring Activity Reports, respectively, issued in March or April after the respective year ends. There was no monitoring program in 2012.

The groundwater monitoring results in 2014 represent the last year of post-construction monitoring. The only construction activities carried out in 2014 were the Inlet Structure Modifications, which did not have potential to affect groundwater.

During the 2014 spring floodway event the Red River water began to flow naturally into the Floodway Channel on April 19, 2014 and continued until April 27, 2014, for a duration of 9 days. The floodway gates were not operated. The Red River crested on April 21, 2014 with a flow of 142 m³/s diverted into the Floodway Channel. The last Floodway Operation with a comparable flow was the 2007 spring flood (119 m³/s) with Floodway Operation. The last operation with a similar duration was in 1998 (8 days).

During the summer flood event the Red River water began to flow naturally into the Floodway Channel on June 30, 2014 and the floodway gates were operated from June 30, 2014 to July 13, 2014 for a duration of 14 days. The Red River crested on July 5, 2014 with a flow of 68 m³/s diverted into the Floodway Channel. The last operation with a similar flow was summer 2011 (64 m³/s). The last summer flood with a similar duration was in 2010 (18 days).

The objective of the Post-Construction monitoring program is to verify the effects predicted in the Project Environmental Impact Assessment and confirm the findings of the groundwater monitoring program during construction, relative to long-term analysis of the groundwater regime.



The requirements for Post-Construction Monitoring Program A were initiated in spring 2014. Spring monitoring is required event if the floodway gates are not operated. Monitoring was not required for the summer flood because he flow and duration was below the specified cut-off (400 m³/s) and duration (8 weeks), however surface water monitoring was initiated because of uncertainties in the flood forecast due to high flows on the Assiniboine River.

Three events were monitored in 2014:

- March 2014 pre-spring run-off (no Floodway Operation) (5 instrumented monitoring wells).
- April 2014 floodway flow with no operation (monitoring wells and surface water monitoring). And May 2014- Post Floodway operation (5 instrumented monitoring wells)
- July 2014 summer Floodway Operation

The carbonate aquifer found along the Floodway Channel is part of a regional groundwater flow system from eastern Manitoba. The confined carbonate bedrock aquifer has natural variations in water quality, with the conductivity ranging from moderate to high (1,000 to 2,000 μ S/cm). Near the Floodway Inlet, local mixing with saline groundwater found west of the Red River results in higher conductivity groundwater (greater than 3,000 μ S/cm) with increased chloride and sodium. Conductivity is a measure of dissolved solids, such as calcium, magnesium, chloride, sodium and sulphate.

Lower conductivity values are found in the bedrock aquifer where it is influenced by the Birds Hill surficial granular aquifer, from CPR Keewatin Bridge to Church Road. The Birds Hill sand and gravel surficial aquifer is a local unconfined aquifer near PTH 59N Bridge. The bedrock aquifer beneath and surrounding the Birds Hill deposit has lower groundwater conductivity due to the freshwater recharge through the sand and gravel.

Natural variations in groundwater quality by location and with the seasons must be considered when the baseline and ongoing water quality results are evaluated during construction activities and Floodway Operation events. One way to detect whether there is surface water intrusion is to monitor an indicator parameter such as conductivity. In the vicinity of the Bird's Hill sand and gravel surficial aquifer, recharge from precipitation results in groundwater with lower conductivity $(500 \,\mu\text{S/cm} \, to 1,000 \,\mu\text{S/cm})$ than is found in other areas of the carbonate aquifer.



The intrusion of any surface water into the groundwater is most readily detected when there is a contrast between the chemistry of the samples. Conductivity (along with other parameters) can be used to evaluate this contrast. Most groundwater conductivity values were found to be greater than surface water conductivity values measured during annual spring Floodway Operation. Red River conductivity values are historically lowest during spring flood events, such as in the spring of 2005, 2006, 2007, 2009, 2010, 2011 and 2013. In this situation, groundwater conductivity would be expected to decrease, if surface water intruded. During summer Floodway Operation in 2005, summer floodway use in 2007, and summer Floodway Operation in 2010 and 2011 conductivity values of water from the Red River diverted in the floodway were slightly higher than in the spring, and higher than the natural groundwater in some areas near the CPR Keewatin Bridge, PTH 59N Bridge and Church Road. These areas have naturally low conductivity. Floodway Channel surface water conductivity was also higher during the summer precipitation events in June 2008 than during the spring melt with no Floodway Operation in April 2008. An increase in groundwater conductivity might occur in summer if surface water intrudes into the groundwater at this time.

In the spring 2014 flood, weekly sampling in the Floodway Channel during the 9-day floodway flow with no operation showed that the conductivity of the surface water increased, with the lowest conductivity found at the beginning of the monitoring. Interpreting changes in groundwater water quality required consideration of the pre-flood conductivity and the changing chemistry of the potential surface water infiltration source.



2.0 POST-CONSTRUCTION MONITORING

Groundwater monitoring in 2010 was considered the first year of the 5-year post-construction monitoring since most channel excavation had been completed. Groundwater depressurization activities in 2010 were completed prior to the spring flood. The 2014 program represents the fifth and last year of the 5-year post-construction period. No further monitoring is required for post-construction effects. Long-term monitoring begins in 2015 focused on key flood events in future years.

In 2014 surface water monitoring of the Floodway Channel during Floodway Operation was carried out at locations near the PTH 44 Bridge and PTH 59N Bridge.

The 2014 post-construction monitoring program used wells designated in the monitoring program for Program A as shown on Figure HM100-1. In the spring 2014 floodway flow with no operation, samples were collected within the Floodway Channel Right-of-Way from 11 bedrock monitoring wells, and one domestic well at the Floodway Inlet structure. Monitoring wells are not used for water supply. Samples for bacteria analysis were taken in five of these monitoring wells in 2014. A monitoring well disinfection program was conducted in fall 2013, prior to sampling in these wells.

3.0 SURFACE WATER RESULTS AND ASSESSMENT

The cold temperature of the Red River at the beginning of the spring melt is generally useful as an indicator of surface water infiltration when the river temperature is a few degrees above freezing and the groundwater temperature is higher. During the 2014 spring flood, the surface water temperature was 2.6 to 4.5°C within the first 2 days of flow. During the summer Floodway Operation in July 2014 surface water temperatures ranged from 17 to 24 °C.

Total Coliform and *E. coli* are present in the surface water. In the 2014 spring flood, Total Coliform bacteria in surface water were present in the 510 CFU/100 mL to overgrown range during most of the flood. Values of *E. coli* ranged from 210 to 1360 CFU/100 mL. During the summer 2014 Floodway Operation, Total Coliform bacteria were greater than 2000 MPN/100 mL. The *E. coli* ranged from 90 to 1180 MPN/100 mL at PTH 59 N.

The maximum concentration of nitrate plus nitrite as nitrogen (N) measured was approximately 1.1 mg/L on April 21, 2013 and was similar 2 days later. The maximum concentration measured in the river water was below the CCME criteria for nitrate plus nitrite (as N) in drinking water (10 mg/L); however, background groundwater concentrations are generally below 0.1 mg/L in many areas of the floodway. Sampling during the summer flood showed maximum nitrate plus nitrite (as N) values of 0.87 mg/L on July 2 at PTH 59 N, dropping to less than detection by July 9, 2014. Ammonia values during the spring flow ranged from 0.14 to 0.18 mg/L, higher than previous values of 0.01 to 0.095 mg/L in 2011 and 2013. Ammonia values during the summer flood were slightly above, or below the detection limit of <0.01 mg/L.

Surface water conductivity increased during the flood, from 286 μ S/cm on April 21, 2014 to 490 μ S/cm on April 25, 2014, reflecting the increase in dissolved solids, hardness, alkalinity, sodium, chloride and sulphate. There was a generally stable water type with a minor change due to an increase in chloride or sulphate concentration. Conductivity increased throughout the 2014 summer flood from 591 μ S/cm on July 2, 2014 to 769 μ S/cm on July 9, 2014 (PTH 59 N). This increase reflects the increase in total dissolved solids, hardness, alkalinity, sodium, chloride and sulphate. Conductivity and major ions increased as nitrate plus nitrite (as N) decreased during the Floodway Operation period.

4.0 GROUNDWATER RESULTS

4.1 FLOODWAY OUTLET AND PTH 44 BRIDGE

At the Floodway Outlet, monitoring wells located 65 m, 160 m and 350 m north of the expanded channel within the Right-of-Way, showed evidence of surface water intrusion in 2014. Changes in conductivity at the Outlet wells corresponded to changes in major ion concentrations in 2014, consistent with previous years. Changes in bacteria (Total Coliform) at the closest monitoring well correlated with inorganic groundwater quality changes in that well. These changes included a decrease in nitrate plus nitrite (as N) concentrations from locally elevated background values. This well closest to the floodway, has the strongest evidence for surface water infiltration, with the intermediate distance well also having similar evidence.

The monitoring well on Rockhaven Road installed in 2013 shows slight evidence of surface water intrusion including decreases in conductivity and major ions, decreases in nitrates from locally elevated background values, and bacteria detection during the peak floodway flow in 2014. These changes may be influenced by the floodway with contributions from other sources including recharge to the aquifer, septic influences and the Red River.

4.2 HAY ROAD TO DUNNING ROAD

Water quality changes were not seen at Hay Road in 2014. There was a small shift in parameters seen at Church Road in spring 2014 that did not correlate with any increase in nitrate plus nitrite (as N) or increase in dissolved oxygen. Water quality was stable at Ludwick Road, Dunning Road and Bray Road.

4.3 PTH 59N BRIDGE AREA

Short-term infiltration of floodway surface water into the bedrock aquifer was documented in 2014 at the PTH 59N Bridge (west side) within the Floodway Right-of-Way. This occurred at a bedrock well located 60 m west of the west channel slope, and a bedrock well 250 m west of the expanded channel at the west Right-of-Way boundary. These responses were consistent with monitoring in previous years, and included changes in conductivity, increasing nitrate plus nitrite



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as (N) and the presence of *E. coli*. Any size or duration operation is likely to cause a short-term response at this location due to the potential for surface water to infiltrate through the sand and gravel in the channel, to the bedrock.

Groundwater quality changes occurred concurrently with water level changes at this location with little time lag. The groundwater quality returned to a typical pre-melt groundwater composition by the time the spring run-off flow in the Floodway Channel had discharged from the channel, and flow was back to the Low Flow Channel (LFC) water level.

The response in this area may represent an initial local recharge through the sand and gravel overburden, followed by infiltration of floodway surface water.

At a monitoring well installed in 2013 at McGregor Farm Road, no change in water quality was seen in 2014.

4.4 KEEWATIN BRIDGE AND AREA TO THE SOUTH

At the Floodway Inlet, the water quality change seen in 2014 appears to be due to surface water infiltration, but was less pronounced in 2014 in comparison to previous years. The lack of change in nitrate plus nitrite (as N) suggests movement of fresher groundwater into the well due to the movement of the saline groundwater boundary to the west during the flood.

5.0 GROUNDWATER RESULTS AND ASSESSMENT

5.1 CONDUCTIVITY CHANGES

Conductivity changes are being used as an indicator of surface water influence on groundwater quality, as conductivity is a parameter that is readily measured. Conductivity changes reflect the changes in major ions contributing to the dissolved solids.

If surface water intrudes into the aquifer, the mixing would result in changes observed in groundwater conductivity. Conductivity decreases with the addition of surface waters in most areas. Changes are most readily observed in areas where groundwater is more mineralized and thus has higher conductivity than surface water, which is typically the case during the spring. Conversely, increases in groundwater conductivity would be seen in areas where baseline groundwater conductivity is less than that of surface waters.

In addition to the floodway, potential surface water infiltration sources in the area include ponds and open sand and gravel quarries, creeks, and the Red River (primarily near the Floodway Outlet).

The magnitude of the water quality change is described by a range in the percentage change in conductivity as follows: Type A (>50% change); Type B (25 to 50% change); Type C (10 to 25% change); Type D (5 to 10% change).

In 2014, 11 monitoring wells and one domestic well (Inlet Control) were sampled for water chemistry. Five of these wells were sampled for bacteria. During spring monitoring in 2014, no obvious change in groundwater quality was seen in 50% (6) of the 12 wells sampled.

Changes found in the other wells were classified as follows:

- Type C (10 to 25% change) for 3 wells (25% of the total) located inside of the Right-of-Way at McGregor Farm Road, the Inlet, Church Road and the Outlet.
- Type B (25 to 50% change) for 3 wells (25%) which were located inside of the Right-of-Way at the Outlet and the PTH 59N Bridge (2 wells).



The wells selected for the sampling in 2014 programs were in areas with higher potential for surface water intrusion due to hydrogeologic conditions, or locations near other surface water sources (such as the Floodway Outlet). In 2014, well sampling occurred north of TCH-1, except for the Floodway Inlet. All of the monitoring wells with water quality changes are located within the Floodway Right-of-Way from north of PTH15 to the Floodway Outlet, except at the Floodway Inlet.

5.2 BACTERIA

Total coliform bacteria were detected in four of the five wells sampled for bacteria at the peak of the flood. The fifth well had no total coliform in 2014. *E. coli* was detected in four of the five wells at the peak of the flood, decreasing to zero in the next sample taken two to three weeks after the flood. Counts of bacteria were higher in 2014 than 2013.

5.3 NITRATE PLUS NITRITE AS (NITROGEN)

Nitrate plus nitrite (as N) concentrations were below the Canadian Drinking Water Quality Guidelines (CDWQG) of 10 mg/L. Nine of the 12 samples, (75%) were below detection (<0.001 to <0.005 mg/L) both pre-melt and during the flood. One sample showed increases in nitrate plus nitrite (as N), where surface water concentrations were greater than groundwater concentrations. One sample showed decreases in nitrate plus nitrite (as N), where surface water concentrations were less than groundwater.

5.4 RELATIONSHIP BETWEEN PARAMETERS

Four of five monitoring wells (80%) sampled for bacteria showed changes in conductivity, Total Coliform and *E. coli* at the peak of the flood. These wells were located inside of the Right-of-Way at PTH 59N Bridge at the west Right-of-Way and channel and inside of the Right-of-Way at two wells at the Floodway Outlet. Bacteria were sampled in areas where changes were expected and locations do not represent a random sample.



Changes in nitrate plus nitrite (as N) correlated with changes in water quality at 33% (5) of 15 wells sampled. Since the maximum nitrate plus nitrite (as N) values in surface water were 0.9 mg/L in 2013, any higher values would be unrelated to Floodway Operation.

Changes in nitrate plus nitrite (as N) correlate with changes in water quality at the Floodway Inlet (one well increasing concentration); PTH 59N Bridge west side (2 wells increasing concentration), and the Floodway Outlet (1 well decreasing concentration). Nitrate plus nitrite (as N) values changed in 4 of the 7 wells where decreases in conductivity (greater than 5%) occurred in spring 2013. The increases in nitrate plus nitrite (as N) were less than 1 mg/L with total nitrate plus nitrite (as N) of 2.0 mg/L or less. Nitrate plus nitrite (as N) concentrations in these wells were below the Canadian Drinking Water Quality Guideline of 10 mg/L nitrate plus nitrite (as N).

Many of the monitoring wells are located on the shoulder of the Floodway Channel, or in the spoil pile, and would be expected to experience any water quality changes more quickly than domestic wells located further away, beyond the Floodway Right-of Way. Domestic wells (with the exception of the Floodway Inlet well) are not monitored in the Post-Construction program. Travel times from the floodway surface water to the monitoring wells vary depending on Floodway Channel water elevations, piezometric water elevations and the hydraulic conductivity of the bedrock, which ranges from highly fractured to massive. In general; however, groundwater gradients will be greater and their travel times will be shorter closer to the floodway. Gradients will decrease and travel times will lengthen further from the floodway.

In cases where conductivity changes appeared to be correlated to floodway use in the spring, the maximum change correlated with conditions of peak flow and surface water elevation during the Floodway Operation period. Water quality started returning to typical pre-spring melt groundwater concentrations as soon as the peak flow started to drop. Water quality returned to pre-melt conditions soon after Floodway Operation ended.

5.5 SUMMARY ASSESSMENT OF CHANGES

The 2014 flood represented a year with a short duration (9 days) and a low discharge (142 m³/s) of Red River water into the floodway with no Floodway Operation. Groundwater quality changes in 2014 were similar to 2013 (a moderate flood year) in some areas and showed less response; in other areas, while bacteria counts were greater than in any year measured 2011 to present. All wells which showed groundwater quality changes in 2014 also showed changes in previous floods. Inorganic groundwater quality changes seen in monitoring wells in 2014 did not exceed the Canadian Water Quality Guidelines for Drinking Water.

Total coliform and *E.coli* bacteria were present in four monitoring wells during the peak of the flood, in areas with documented groundwater/surface water connections, with concentrations decreasing to zero for *E. coli* and zero to trace concentrations for total coliform after Floodway Operation.

Bacteria counts of Total Coliform and *E. coli* increased above previously monitored values in instrumented monitoring wells in the sensitive areas that were identified as zones of potential surface water infiltration. Because of this, an adverse effect on groundwater may have occurred for a short duration and these areas may have been at greater risk of surface water infiltration than in previous years. Special reporting to Manitoba Water Stewardship/Manitoba Conservation was required and conducted, indicating that these specific areas may have been at greater risk of surface water infiltration than in previous years, for a short duration during the spring flood.

6.0 SPRING TREATMENT AREAS

The Spring Treatment Program has mitigated surface water infiltration in the bedrock aquifer by providing sand filtration of any fines migration, by decreasing the amount of flow into the springs at the filter locations for a given flood, by reducing the potential for expansion of spring areas through piping, and by improving the bacterial quality of any infiltrating water. The constructed fine sand filters have a much lower hydraulic conductivity than an open fracture; therefore, the initial flow rate is decreased. As the low permeability silt fraction builds up above the sand filter layer, the infiltration rate is reduced further. The fine sand also meets criteria for slow sand filters designed to reduce bacteria passage through the filter. After the flood, when the flow direction reverses to groundwater discharge, the sand filter protects against upward piping of the foundation material (silt, sand) which otherwise could have increased the size of a fracture/hole.

Sealing the groundwater discharge areas completely is not desirable, as a pressure build-up and uncontrolled discharge in another area would likely develop. The treatments provide pressure relief, but in a controlled fashion and with a flow rate lower than was present before treatment.

Previous sampling in 2009 through 2011 showed that Total Coliform and *E. Coli* bacteria are generally present and at higher levels above the filter. The filter has been effectively reducing Total Coliform concentrations. *E. coli* has not been detected beneath the filter. Soon after the floodway drains, surface water infiltration is flushed out quickly from the system as shown by a return to groundwater quality and an absence of bacteria. A return to groundwater quality (as shown by conductivity) was seen towards the end of the flood period as shown by the transducer data in 2009, 2011 and 2013.

In 2014 one spring discharge location was monitored with a transducer. The results from 2014 were similar to 2013, 2011 and 2009, showing that floodway surface water temporarily can reverse the gradient and flow direction from the channel into the bedrock beneath the spring, with a return to groundwater quality as the floodway begins to drain and groundwater is again discharged into the channel.

A summer inspection of 23 spring locations treated with reverse sand filters was conducted in 2014 to update the 2013 survey. Sketches, photographs and video were collected at the same locations as 2013 to compare conditions and can be used in the future to assess if there are changes in conditions. Filters are operating as designed and no new discharge areas were observed.



7.0 GROUNDWATER ACTION RESPONSE PLAN AND LONG-TERM MONITORING

The Groundwater Action Response Plan has been continued through the Post-Construction period. No complaints were received in 2014.

Monitoring in the last two years of the post-construction monitoring program supports the conclusions reached in previous reports of no adverse effect of the floodway expansion on surface and groundwater quality. Changes in the long-term monitoring program are needed to streamline monitoring for small flood events and provide clear program direction for a wider range of operating and non-operating conditions. A revised long-term monitoring program has been proposed beginning 2015 designed to accommodate differing spring and summer flood conditions including peak flows and durations.

8.0 STATEMENT OF LIMITATIONS AND CONDITIONS

8.1 THIRD PARTY USE OF REPORT

This report has been prepared for the Manitoba Floodway Authority to whom this report has been addressed and any use a third party makes of this report, or any reliance on or decisions made based on it, are the responsibility of such third parties. KGS Group accepts no responsibility for damages, if any, suffered by any third party as a result of decisions made or actions undertaken based on this report.

8.2 GEO-ENVIRONMENTAL STATEMENT OF LIMITATIONS

KGS Group prepared the geo-environmental conclusions and recommendations for this report in a professional manner using the degree of skill and care exercised for similar projects under similar conditions by reputable and competent environmental consultants. The information contained in this report is based on the information that was made available to KGS Group during the investigation and upon the services described, which were performed within the time and budgetary requirements of the Manitoba Floodway Authority. As the report is based on the available information, some of its conclusions could be different if the information upon which it is based is determined to be false, inaccurate or contradicted by additional information. KGS Group makes no representation concerning the legal significance of its findings or the value of the property investigated.

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FIGURES

